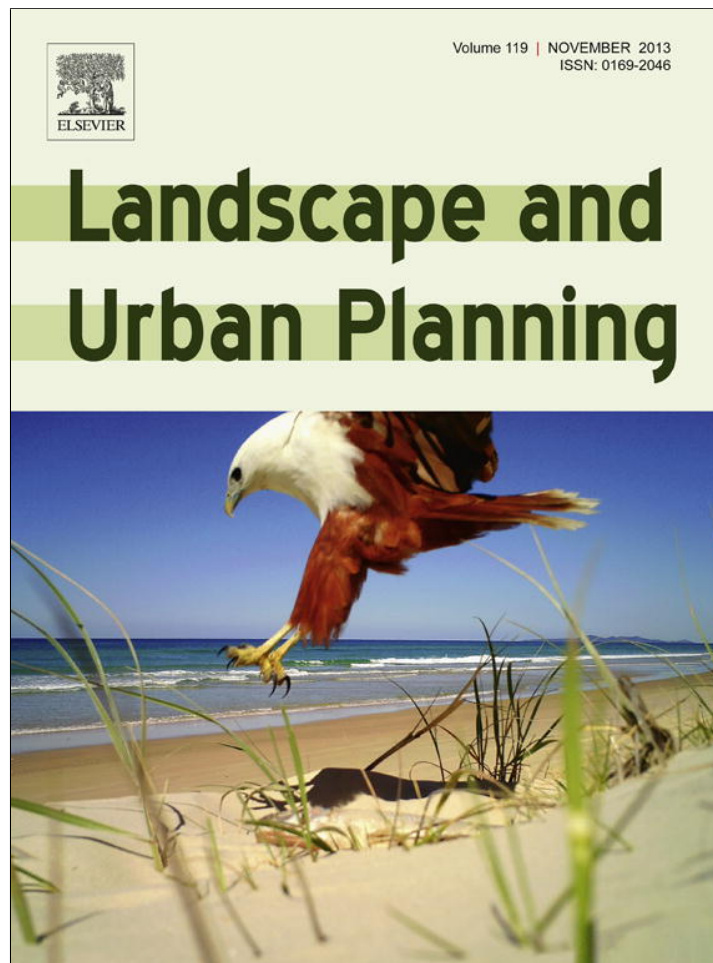


Provided for non-commercial research and education use.
Not for reproduction, distribution or commercial use.



This article appeared in a journal published by Elsevier. The attached copy is furnished to the author for internal non-commercial research and education use, including for instruction at the authors institution and sharing with colleagues.

Other uses, including reproduction and distribution, or selling or licensing copies, or posting to personal, institutional or third party websites are prohibited.

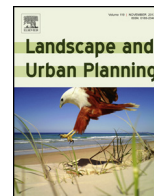
In most cases authors are permitted to post their version of the article (e.g. in Word or Tex form) to their personal website or institutional repository. Authors requiring further information regarding Elsevier's archiving and manuscript policies are encouraged to visit:

<http://www.elsevier.com/authorsrights>



Contents lists available at ScienceDirect

Landscape and Urban Planning

journal homepage: www.elsevier.com/locate/landurbplan

Research paper

Urbanisation alters processing of marine carrion on sandy beaches

Chantal M. Huijbers^{a,b,*}, Thomas A. Schlacher^b, Dave S. Schoeman^b, Michael A. Weston^c, Rod M. Connolly^a^a Australian Rivers Institute – Coast & Estuaries, and School of Environment, Griffith University, Gold Coast, Qld 4222, Australia^b Faculty of Science, University of the Sunshine Coast, Maroochydore DC, Qld 4558, Australia^c Deakin University, Centre for Integrative Ecology, School of Life and Environmental Sciences, Deakin University, Burwood, Victoria 3125, Australia

HIGHLIGHTS

- Coastal urbanisation significantly alters the composition of scavenger assemblages.
- Invasive mammals replace iconic raptors as scavengers on urban sandy beaches.
- Raptors rapidly remove marine carrion from rural beaches.
- Profound effects of land-use changes on key ecological processes.

ARTICLE INFO

Article history:

Received 4 February 2013

Received in revised form 13 June 2013

Accepted 13 June 2013

Available online 21 July 2013

Keywords:

Scavenging

Shoreline management

Human impacts

Food web subsidies

Raptors

ABSTRACT

Sandy shores are highly attractive for urban development. Urbanisation of beaches is, however, not without environmental consequences, but metrics of ecological change along metropolitan coasts are poorly developed. This lack of metrics impedes environmentally effective coastal zone management. Here we test the effects of urbanisation on a pivotal ecological process on sandy shorelines: carrion removal by vertebrate scavengers. Scavenging is a key process linking ocean and land ecosystems via animal carcasses deposited on beaches and subsequently consumed by mostly terrestrial animals. In this study, experimentally placed fish carcasses were monitored with motion-triggered cameras on urban and rural beaches on the east coast of Australia. Urbanisation substantially influenced the structure of the scavenger guild and the frequency of carrion removal within 24 h. Large raptors were abundant on rural beaches where they rapidly detected and consumed carrion (98% of carcasses removed within 24 h). We detected no scavenging activity of raptors on urban beaches, where scavenging birds of prey were functionally replaced by nocturnally foraging, non-native mammals (red fox, *Vulpes vulpes*) or feral species (cats, dogs) known to threaten beach-dwelling wildlife. Our findings emphasise the value of non-urbanised coastal dunes and sandy beaches as important feeding sites and habitats for iconic and threatened raptors. We also show that human changes in coastal land-use profoundly alter ecological structures and processes on sandy shorelines, aspects that warrant explicit inclusion in landscape management and planning of the coastal strip.

Crown Copyright © 2013 Published by Elsevier B.V. All rights reserved.

1. Introduction

Coastal cities are hotspots of environmental change, driven by the continual expansion of the urban footprint in a narrow strip of land bordering the oceans (Martínez et al., 2007). This growth of coastal urbanisation has serious environmental consequences,

such as the loss and fragmentation of habitats, overexploitation of fish stocks, pollution, and depletion of populations of coastal species (Lotze et al., 2006; Schlacher & Thompson, 2012). Environmental management of the growing urban footprint in coastal areas is a formidable challenge, chiefly because the main objectives of protecting infrastructure and maximising amenity values often dominate the agenda at the cost of environmental conservation (Duxbury & Dickinson, 2007; Schlacher et al., 2006, 2007).

Sandy beaches geographically dominate coastlines around the world, and much of the coastal urban development occurs behind sandy beaches (Gurran, 2008; Noriega, Schlacher, & Smeuninx, 2012). Beaches are the longest ecological interface between the oceans and the land, forming hotspots of bio-geochemical processing of material (Schlacher et al., 2008). Almost all of the

* Corresponding author at: School of Environment, Griffith University, Gold Coast campus, Qld 4222, Australia. Tel.: +61 7 55529189.

E-mail addresses: c.huijbers@griffith.edu.au (C.M. Huijbers), TSchlach@usc.edu.au (T.A. Schlacher), dschoema@usc.edu.au (D.S. Schoeman), mike.weston@deakin.edu.au (M.A. Weston), r.connolly@griffith.edu.au (R.M. Connolly).



Fig. 1. (A) Location of the study area in Eastern Australia, with (B) exact locations of three sections of beach along the rural Noosa North Shore (left panel), and three urban beaches in the vicinity of the town of Mooloolaba (right panel).

material processed in beach ecosystems is imported from the sea, making beaches prime examples of 'subsidised' food webs (Polis & Hurd, 1996). Carrion (animal carcasses, usually washed ashore) is a distinct source of marine subsidy to beaches, and a diverse suite of scavengers utilise this resource (McLachlan & Brown, 2006). Scavengers on beaches tend to be facultative rather than obligate scavengers, and are little studied (Beasley, Olson, & DeVault, 2012).

Scavenging, on beaches and elsewhere, is a key ecological process (Barton, Cunningham, Lindenmayer, & Manning, 2013). Numerous factors (e.g., carrion availability, predictability, quality) influence the diversity and type of scavengers, and the rate at which carrion is processed (DeVault, Rhodes, & Shivik, 2003). Notwithstanding this variability, scavenging is seen as a pivotal pathway of energy transfer in most food webs (Wilson & Wolkovich, 2011). Previously, most research has focused on scavenging communities in undisturbed habitats (e.g., Ogada, Torchin, Kinnaird, & Ezenwa, 2012; Parmenter & MacMahon, 2009; Selva, Jędrzejewska, Jędrzejewski, & Wajrak, 2005), with only a few studies addressing this important ecological process in ecosystems affected by human impacts (DeVault, Olson, Beasley, & Rhodes, 2011). Because sandy beaches are systems that are underpinned energetically by imports of organic matter (overwhelmingly from the sea), scavenging is predicted to be a key functional element of beach systems (Rose & Polis, 1998; Schlacher, Strydom, & Connolly, 2013), yet how this process is influenced by urbanisation of sandy beaches is unknown.

We used a field experiment to assess the effects of coastal urbanisation on scavengers and necromass consumption on sandy beaches. Urbanisation globally causes biotic homogenisation, often at the cost of native species (McKinney, 2006), and thus is predicted to alter the species richness and composition of scavenger guilds on urban beaches. We specifically tested if: (a) the species composition of the scavenger guild at beaches and dunes differs between urban and rural areas, and (b) whether such structural changes have consequences for the short-term scavenging efficiency (i.e., carcass removal within 24 h) of carrion from the shoreline.

2. Methods

2.1. Study area

This study was conducted on the Sunshine Coast in southeast Queensland, Australia (Fig. 1A). This area is one of the fastest-growing coastal regions in Australia, and is intensively used for beach recreation. Areas of intense coastal urbanisation, such as the town of Mooloolaba, are interspersed with rural coastal areas, such as those found to the north of the Noosa River Estuary (Fig. 1B). We use the term 'rural' as defined by the Australian Bureau of Statistics as being non-urban, having extremely low levels of human habitation and generally natural vegetation.

The coastline of Mooloolaba has been significantly transformed over the past century by urban development, where infrastructure (e.g., boardwalks, seawalls, playgrounds, roads, buildings, etc.) has largely replaced natural ecosystems. This has reduced dune width, and completely eliminated dunes in some places (Longhurst, 1997; Lucrezi, Schlacher, & Walker, 2009). By contrast, beaches and dunes north of the Noosa Estuary are located within the Great Sandy National Park where infrastructure is minimal and extensive dunes back the 50 km of continuous beach.

2.2. Experiments

We measured species composition of the scavenging guild and short-term carcass removal rates on three urban and three rural beaches (Fig. 1). The urban beaches were centred on the town of Mooloolaba (Kawana Beach, Mooloolaba Beach, and Alexandra Headland; 26.40° S, 153.07° E), while the rural beach sections were located on the Noosa North Shore (26.15° S, 153.05° E).

In our experimental design, 'Urbanisation' represents a fixed factor with two levels: urban and rural. 'Beach' is a random factor with six levels, nested within 'Urbanisation'. Each beach contained four replicate sites for experimental observations. The distance between beaches, within treatments, was ~2 km, while replicate sites within beaches were spaced ~200 m apart along the shore. The experiment was repeated five times over a six-week period in June–August 2012, adding a final random effect ('Trial') to our design. The average minimum and maximum temperatures (\pm s.d.) for the five days on which sampling occurred were 8.0 (\pm 3.6) °C and 21.3 (\pm 1.9) °C, respectively (data from Bureau of Meteorology, Australia). In each trial, carrion was deployed at the same sites. In total, we conducted 120 camera deployments, but retrieved results from only 114 cameras due to theft or interference ($n=59$ for rural beaches, $n=55$ for urban beaches). Due to logistical constraints, trials on urban and rural beaches were conducted on separate, but consecutive days. The starting time of deployments varied haphazardly among trials, but was matched between urban and rural beaches within a trial. To ensure that both diurnal and nocturnal scavengers could be detected if present, we deployed carrion during the day (between 10 am and 3 pm) in trials 1, 2 and 4, and just before sunset (5 pm) in trials 3 and 5. Deployments were on average 12 days apart. To ascertain that our deployments did not lead to learned-attraction to camera locations of the scavengers, we fit linear mixed-effects models to investigate correlations with time to scavenging (in min) for daytime and sunset deployments separately. For sunset deployments, Trial ($p=0.595$), Urbanisation ($p=0.161$) and their interaction ($p=0.297$) were all non-significant, indicating that there was no discernible learning effect at night. For daytime deployments, time to scavenging was significantly faster for Trial 2 than for either Trial 1 ($p=0.007$) or Trial 4 ($p=0.038$), although again the Urbanisation effect ($p=0.309$) and the interaction between Trial and Urbanisation ($p=0.824$) were non-significant. These results suggest that there is no evidence that scavengers systematically improved their ability to detect the deployed carcasses.

In each camera deployment, one fish carcass (383 ± 113 g) was placed on the beach near the dunes. Flathead grey mullet, *Mugil cephalus*, was used as carrion because this species is commonly found in the surf-zone of beaches worldwide, and as carrion is scavenged by several species (Schlacher et al., 2013). Each fish was monitored for 24 h with a digital passive infrared (PIR), motion-triggered camera (ScoutGuard SG560Z-8M). This technique is commonly used to assess the occurrence and nature of scavenging at carcasses (Hamel et al., 2012). Each camera was set to take three consecutive pictures upon detection of movement, with a reset period of 7 s in between trigger events. The cameras took 136,420 pictures in total, ranging from 15 to 5508 images per deployment. Large numbers of images were not necessarily the result of large

numbers of scavengers because other moving objects also triggered cameras (e.g., vegetation moved by the wind, walkers, cars, etc.). We elected to expose carrion for a standardised period of 24 h chiefly because urban beaches are periodically groomed and longer exposures would thus have introduced the possibility of carcass removal by beach cleaners, which would consequently lead to incomplete trials. No evidence of grooming was observed on the day of carcass placement during the experiments. We report all results against this standardised exposure time.

Animals were recorded as scavengers when they were detected with the fish in their mouth, with their mouth touching the fish, or when they appeared next to the fish in an image and the fish was missing from the next image.

2.3. Data analysis

The species richness of the scavenger guild was compared between urban and rural areas using sample-based species-accumulation curves, as determined from random permutations of the data, or subsampling without replacement (Gotelli & Colwell, 2001). In estimating species richness, the number of scavenger species (as defined above) recorded at each carrion deployment was recorded, regardless of the number of individuals present to avoid overrepresentation of species that naturally scavenge in groups rather than individually. Assemblage composition by beach and trial was then assessed by aggregating scavenger species incidences across all sites within individual beaches for each trial. Non-metric multidimensional scaling (nMDS; based on Bray–Curtis resemblance matrices calculated from untransformed data of aggregated species incidence per beach and trial) was used to visualise differences in scavenger species composition among beaches and trials. To test whether the structure of the scavenger guild differed between urban and rural beaches (the main test of interest according to our hypothesis), we used a simple additive permutational multivariate analysis of variance (PERMANOVA) to fit a linear model to Bray–Curtis dissimilarity matrices (Anderson, 2001; Oksanen et al., 2012). Three trials from the urban beach 'Alexandra Headland' recorded zero scavengers and therefore had to be omitted from PERMANOVA, making the test a conservative one (i.e., differences between urban and rural scavenger community composition are larger than suggested by PERMANOVA).

SIMPER analysis (in PRIMER) was used to highlight which species contributed most to the dissimilarity in community composition between urban and rural beaches. Because violations of mean-variance assumptions may confound dispersion and location effects (Warton, Wright, & Wang, 2012), we supplemented SIMPER analysis with Poisson mixed-effects linear models (Bolker et al., 2009) for each species separately to test for species-level effects of urbanisation. In each case, we modelled the number of times a scavenger species was observed as a function of urbanisation (fixed effect), beach (random effect, to account for repeated sampling at specific beaches nested within levels of urbanisation), and individual observation (to account for overdispersion of the Poisson regression). Significance was assessed on the basis of a likelihood-ratio test of the fit of model containing the urbanisation factor relative to that for the model without the urbanisation factor (Bolker et al., 2009; Zuur, Ieno, & Smith, 2007).

The likelihood of a carcass being removed by scavengers within 24 h (binary outcome) was modelled using a logistic mixed-effects model with binomial error structure. Our full model included level of urbanisation as a fixed effect having two levels (urban and rural), with trial and beach as random effects because their levels were selections from larger pools to which we wish to generalise our results. Generalised linear mixed-effects modelling is powerful not only because it models data using a realistic error structure, but also because it accommodates missing values (cameras were not

Table 1
The frequency of occurrence of scavenger species on urban and rural beaches, determined as the number of camera deployments in which a species was positively recorded consuming carrion. Species are ordered according to their percentage contribution (based on SIMPER) to the dissimilarity in community composition between urban and rural beaches. 'Undetermined' denotes incidences where fish carcasses were removed from the beach but the actual scavenger species consuming the carrion was not captured by the camera, most likely a bird of prey swooping on fish carcasses and lifting them off the beach (Schlacher pers. obs.). '-' accounts for species that contributed less than 10% to the dissimilarity between urban and rural beaches. Mixed-effects model *p* values indicate the significance (according to a likelihood-ratio test) of urbanisation for Poisson generalised mixed-effects linear models of scavenging occurrences for each of the species, separately (see Section 2).

| Scavenger species | Occurrence | | % contribution to dissimilarity | Mixed-effects model <i>p</i> |
|-------------------------|------------------------|------------------------|---------------------------------|------------------------------|
| | Urban (<i>n</i> = 55) | Rural (<i>n</i> = 59) | | |
| Brahminy kite | 0 | 33 | 36.7 | <0.0001 |
| Undetermined | 3 | 15 | 15.0 | 0.0235 |
| Whistling kite | 0 | 11 | 12.2 | 0.0002 |
| Red fox | 10 | 3 | 10.8 | 0.8343 |
| Torresian crow | 4 | 6 | 8.5 | 0.5809 |
| Rat | 7 | 0 | 7.9 | 0.0262 |
| Domestic dog | 4 | 0 | – | 0.3730 |
| Domestic cat | 2 | 0 | – | 0.2100 |
| Silver gull | 2 | 0 | – | 0.1386 |
| White-bellied sea eagle | 0 | 1 | – | 0.5631 |
| Lace monitor | 0 | 1 | – | 0.5631 |

recovered on six occasions), and it specifically acknowledges the nested structure of the survey design, thereby accounting for the fact that we repeatedly revisited the same sites on the same beaches (Bolker et al., 2009; Zuur et al., 2007). The full model was simplified following the suggestions of Zuur et al. (2007) and Bolker et al. (2009), namely: starting with the full model and searching for optimal model structure by sequentially dropping random and then fixed effects. Improved model fit is indicated by a smaller Akaike information criterion (AIC), a smaller Bayesian information criterion (BIC) and a strongly non-significant likelihood-ratio test, which indicates that the fit of the simpler model did not cause a significant deterioration in fit in terms of deviance (–2 times the log-likelihood). Because the likelihood-ratio test can be biased for fixed effects (Bolker et al., 2009), we verified the results in this instance with a Wald χ^2 test.

3. Results

Urbanisation substantially changed the community structure of the scavenger guild and the probability of carrion removal over a 24 h exposure period. Although species richness was comparable between urban and rural beaches (6 species for both levels of urbanisation), the composition of the scavenger guild was significantly different (PERMANOVA, Urbanisation effect: $F_{1,21} = 20.701$, $p = 0.001$).

The large structural separation of the scavenger guild between rural and urban beaches was mainly due to raptors (brahminy kite, *Haliastur indus*, and whistling kite, *H. sphenurus*) being the dominant scavenger species on rural beaches, whereas feral mammals (rats, *Rattus* spp.), were more common on urban beaches (Table 1, Fig. 2). The only scavenger species detected on both rural and urban beaches were red foxes (*Vulpes vulpes*) and Torresian crows (*Corvus orro*). These species were recorded at comparable frequencies on both beach types, but accounted for a significantly larger proportion of necromass removal on urban shores. Scavengers remained unidentified when the fish carcass was no longer present but no scavenger was captured on the images. This occurred 3 times on urban beaches and 15 times on rural beaches, and most likely results from animals moving faster than the trigger delay of the cameras.

Most (64%) of the fish carcasses that we had deployed on urban beaches were still present after 24 h: the few positive records of scavengers on urban beaches consisted mainly of red foxes and rats (*Rattus* spp.), complemented by Torresian crows, domestic dogs (*Canis familiaris*), domestic cats (*Felis catus*), and silver gulls (*Chroicocephalus novaehollandiae*) (Fig. 3). On urban beaches only

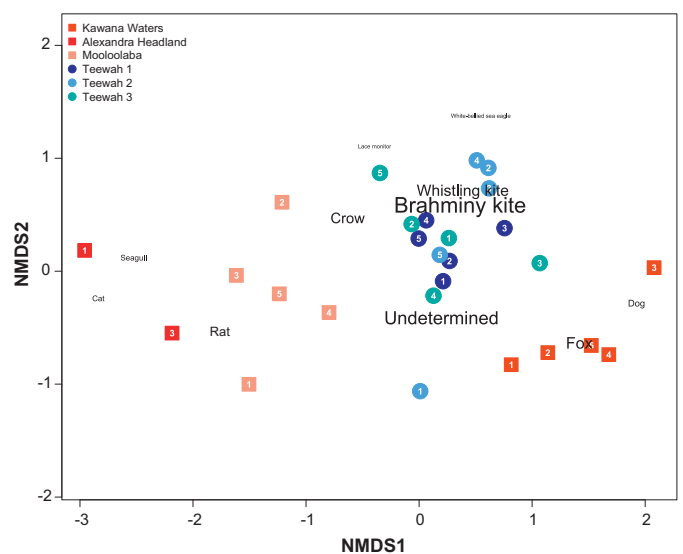


Fig. 2. nMDS biplot (based on Bray–Curtis resemblance from untransformed data) of scavenger species composition at experimental deployments of carrion on urban (squares in shades of red) and rural (circles in shades of blue) beaches (numbers represent different trials). Font size of scavenger names is scaled to reflect the relative frequency of observation for a species, and positioning of names reflects associations with the nMDS ordination. (For interpretation of the references to color in figure legend, the reader is referred to the web version of the article.)

foxes, and on three occasions domestic dogs, removed entire fish carcasses, while the other scavenger species only fed on small proportions of the fish (<20% of the mass). The differences in scavenger guilds were closely related to the time of day at which scavenging occurred (Table 2). Scavenging guilds on urban beaches were dominated by mammals that were mainly active at night, while rural scavengers were mostly avian species that foraged during the day.

Table 2
Comparison of the total number of individual avian and mammalian scavengers, recorded during either day or night time with motion-triggered cameras on urban and rural beaches. The presence of avian and mammalian scavengers is closely related to the time of day when scavenging occurred.

| | Birds | | | Mammals | | |
|-------|-------|-------|-------|---------|-------|-------|
| | Day | Night | Total | Day | Night | Total |
| Urban | 3 | 3 | 6 | 3 | 19 | 22 |
| Rural | 51 | 0 | 51 | 0 | 3 | 3 |
| | 54 | 3 | 57 | 3 | 22 | 25 |



Fig. 3. Examples of scavengers detected on rural and urban beaches: (A) brahminy kite (*Haliastur indus*), (B) whistling kite (*H. sphenurus*), (C) lace monitor (*Varanus varius*), (D) Torresian crow (*Corvus orro*), (E) silver gull (*Chroicocephalus novaehollandiae*), (F) red fox (*Vulpes vulpes*).

Species-accumulation curves for both rural and urban beaches converged on the six identified scavenger species, and although the curve is initially slightly steeper for rural than urban beaches, the asymptote is reached more quickly for urban than rural beaches (Fig. 4). The curve for urban beaches plateaued earlier because the scavenging community of rural beaches contained two singletons (lace monitors *Varanus varius* and white-bellied sea eagles *Haliaeetus leucogaster*), whereas urban beaches had none, suggesting that the structure of the urban scavenging community may have been slightly better quantified by our survey methods than that of the rural beaches.

Scavenging of carrion from rural beaches was rapid and intense: 98% of experimentally deployed fish carcasses were removed by scavengers within 24 h. Initial inspection of the logistic mixed-effects model indicated that the random effect for Trial was non-significant ($2 \times \Delta\text{Log-likelihood} = 0.03$, $df = 1$, $p = 0.863$), but that the random effect for Beach was significant ($2 \times \Delta\text{Log-likelihood} = 25.563$, $df = 1$, $p < 0.0001$). The fixed effect for Urbanisation was strongly significant ($2 \times \Delta\text{Log-likelihood} = 6.5877$, $df = 1$, $p = 0.01$; Wald $\chi^2 = 7.4515$, $df = 1$, $p = 0.006$). Estimates from the final model confirm that the likelihood of a carcass being removed

by scavengers from rural beaches within 24 h was close to 100% (Fig. 5). By contrast, the likelihood of a carcass being removed from an urban beach within 24 h is much lower (22% on average, but with a wide confidence interval), which is no different from even odds.

The time of deployment affected the time lag between deployment and occurrence of a scavenger at a carcass. Daytime deployments led to quick detection by diurnally active avian species on rural beaches: 86% of all fish carcasses that were deployed during the day were detected by scavengers within 4 h of placement on the beach (mean \pm s.d. time to first scavenger: $3:02 \pm 5:52$ h). By contrast, when carrion was placed on rural beaches close to sunset, scavengers were much slower to detect carcasses: 55% of these carcasses remained after 12 h, and many of these were removed the next day by diurnal scavengers (mean \pm s.d. time to first scavenger: $10:03 \pm 8:18$ h). Arrival of scavengers at an experimental carcass was more protracted on urban beaches, both for diurnal and nocturnal scavengers. On urban beaches, the mean time (\pm s.d.) to detection for daytime deployments was $7:32 (\pm 4:59)$ h. The frequent occurrence of foxes, mostly crepuscular and nocturnal foragers, on urban beaches resulted in

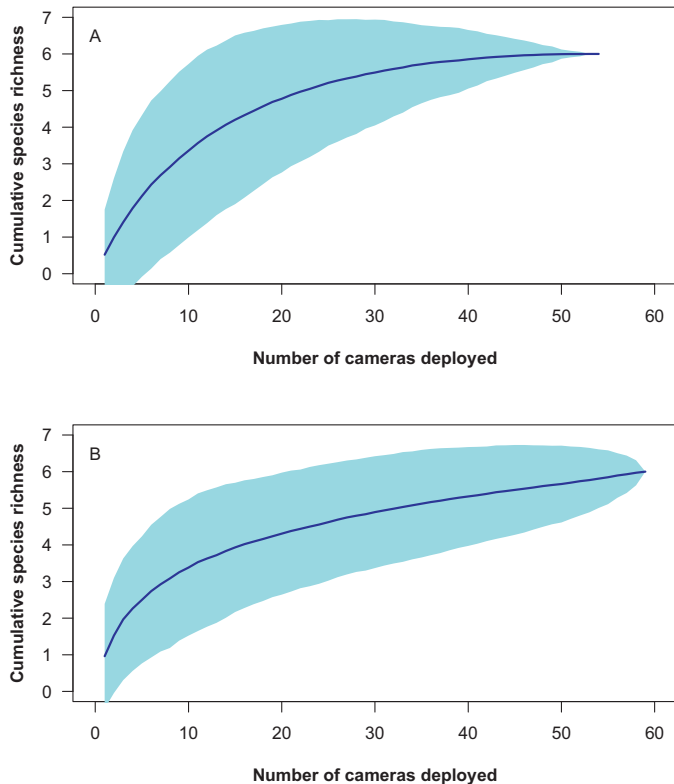


Fig. 4. Species-accumulation curves for scavenger communities observed at experimental carrion deployments at (A) urban, and (B) rural beaches. Blue areas represent the standard deviations as determined from 1000 random permutations of the data (subsampling without replacement). (For interpretation of the references to color in figure legend, the reader is referred to the web version of the article.)

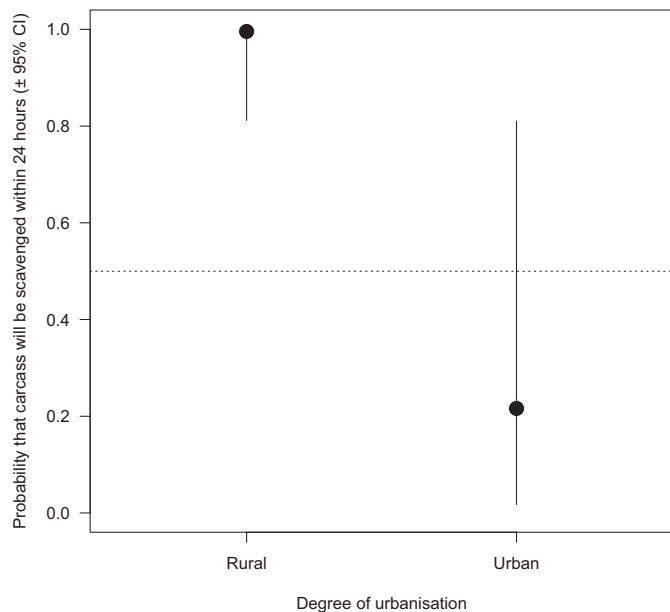


Fig. 5. Probability (calculated from a logistic mixed-effects model with binomial error structure) that a fish carcass will be removed within 24 h from rural and urban beaches. The dashed line represents the 50:50 chance that a carrion item is removed by scavengers within the observation period of 24 h.

relatively quick detection times ($4:06 \pm 4:41$ h) in deployments done close to sunset.

4. Discussion

Urbanisation significantly altered scavenging patterns on sandy beaches of southeast Queensland, Australia: urban beaches supported a suite of scavengers, including non-native and feral mammals, which removed significantly less carrion within 24 h than did the raptor-dominated scavenger communities on rural beaches. Although our study is the first to compare removal of marine carrion by terrestrial and avian vertebrate scavengers simultaneously on urban and rural beaches, our results correspond to those from a previous study of scavenging rates in six tropical habitats along an urbanisation gradient, which showed longer detection times and lower scavenging rates by insects on dead invertebrates in more urbanised areas (Tan & Corlett, 2012). This suggests a generalisation in the ways in which changes in land-use, and especially urbanisation, can affect key ecological processes.

The consequences of such changes to fundamental trophic processes in coastal food-web structure could be profound, and should be taken into consideration by coastal planners and managers. A significant and accelerating threat being faced by the narrow strip of coast that comprises sandy beaches is coastal squeeze (Schlacher et al., 2008). This phenomenon is caused by rising sea levels, which combine with expanding urbanisation to trap and slowly drown beaches. Trends in sea-level rise are impossible to alter in the short to medium term, so to conserve beaches, coastal managers can only regulate urban and peri-urban development. Planning the type and spatial location of development that might irreversibly alter habitat dimensions, habitat connectivity, or quality, requires an understanding of how changes in coastal habitats affect critical ecosystem services. Management of beaches has traditionally focused on maintaining recreational opportunities in urban settings, with scant regard to conserving ecological processes in urban planning of sandy shorelines (Schlacher et al., 2006, 2007). The lack of robust metrics by which to assess the effects of human interventions on beach systems has been considered to impede more ecologically focused approaches to beach management (James, 2000; Micallef & Williams, 2002; Schlacher et al., 2006). Our metric of scavenging captures an essential mechanism of nutrient and organic matter transfer in ecosystems at the land-ocean interface that has potential in environmental monitoring and planning applications. For example, our key finding of the importance of large avian scavengers in the systems studied has implications for conserving and providing critical nesting and foraging sites for these threatened vertebrates.

Dominant scavenger species in rural areas (brahminy and whistling kite) were not detected by us in urban areas. Clergeau, Croci, Jokimäki, Kaisanlahti-Jokimäki and Dinetti (2006) showed that avian species composition in urbanised areas is generally simplified, with a paucity of habitat specialists. Many species of raptor are highly sensitive to urbanisation, which has led to a dramatic decrease in numbers of some raptor species worldwide (Berry, Bock, & Haire, 1998; Eduardo, Carvalho, & Marini, 2007). Furthermore, the scavenging raptors we report are predominantly diurnal feeders, and their absence from urban beaches might be related to human presence and disturbance during the day. Larger bird species exhibit an earlier response to human disturbances, and move further away after being disturbed (Blumstein, Fernandez-Juricic, Zollner, & Garity, 2005; Weston, Mcleod, Blumstein, & Guay, 2012). The avian scavengers we recorded on urban beaches (e.g., gulls and crows) tend to be commensal with, or at least highly tolerant, of human presence (Weston et al., 2012). The fact that we detected no significant scavenging activity of raptors on urban

beaches suggests that densities are too low to be a major influence on carrion processing in urban areas. This highlights the impact of urban development on beach and dune-dependent species that occur naturally in less disturbed areas (Meager, Schlacher, & Nielsen, 2012). Ironically, although we show that coastal development alters the wildlife species occurring on beaches, many coastal residents desire wildlife on beaches (Maguire, Miller, Weston, & Young, 2011).

In addition to altered scavenger guilds, urbanisation also affected the amount of carrion that was removed within 24 h from sandy beaches. All but one of the deployed fish carcasses were removed within 24 h from the beaches in the rural area, compared to only 20 out of 55 on urban beaches. Removal rate is partly influenced by the composition of scavenger species (Bodkin & Jameson, 1991). Large raptors and foxes tend to consume fish carcasses completely, or at least to remove them from the beach, in contrast to smaller mammalian and avian scavengers that dominated on the urban beaches. The raptors reported in our study search for their food with an energetically efficient soaring locomotion (Ruxton & Houston, 2004), resulting in a much quicker and larger removal rate of carrion compared with mammalian scavengers. It is plausible that some fish carcasses might eventually have been removed by scavengers had we left them on the beach beyond 24 h, for example by scavengers that mainly use olfactory cues to detect carrion, or by occasional forays into urban settings by primarily rural raptors. However, urban beaches tend to be regularly groomed by local councils, so such scavengers are unlikely to contribute significantly to coastal energy flows. Therefore, although we acknowledge that deliberate removal of the fish carcasses after 24 h does not reveal the ultimate fate of carrion, our results clearly show that urbanisation has several impacts on the scavenging process: by altering the species composition of the scavenger guild, urbanisation reduces the rate of carrion consumption, which might have broader impacts throughout coastal food webs.

The rapid and intense removal of the fish carcasses on rural beaches equals scavenging rates in reported for some other ecosystems. Below sea level, fish removed all carrion deposited on coral reefs within 24 h (Rassweiler & Rassweiler, 2011). In terrestrial systems, removal rates of >95% within 3 days have been observed in tropical forests (Houston, 1986) and agricultural fields (Peterson, Lee, & Elliott, 2001), where raptors such as vultures, bald eagles and harriers dominated the scavenging guild. Despite the significant and substantial effects of urbanisation on scavengers and scavenging found by us, carrion removal by vertebrate scavengers continues on urbanised beaches (i.e., 36% of the carrion was removed within 24 h). This demonstrates the pervasive ecological role of scavenging even in highly modified systems (Wilson & Wolkovich, 2011). Importantly, the disruption of native scavenging assemblages seems to assist generalist vertebrate pest species, which have been reported to prey on wildlife, such as beach-nesting birds (Maguire, Stojanovic, & Weston, 2010). Furthermore, the presence of introduced species (foxes, feral dogs and cats), that have been repeatedly demonstrated to be serious threats to native wildlife (Loss, Will, & Marra, 2013; Short & Smith, 1994), compounds the negative effects of urbanisation on coastal ecosystems.

Food-web subsidies are critical to the functioning of sandy beaches (Schlacher et al., 2008). The consumption of fish carcasses by terrestrial scavengers forms an important conduit by which marine nutrients are translocated into terrestrial systems (Polis, Sánchez-Piñero, Stapp, Anderson, & Rose, 2004). Although some scavengers consume carrion completely, remains of the fish carcasses can be consumed by species of lower trophic position such as ghost crabs (Wolcott, 1978) or wolf spiders (Morse, 1997). In this way, marine subsidies can propagate across multiple levels in coastal food webs (Spiller et al., 2010). It is clear from our research that these processes differ between areas with different levels of

human development. In rural areas, where avian scavengers dominate the transfer of marine carrion to coastal food webs, marine subsidies are likely to be transferred over much larger distances compared with urban areas with small mammalian scavengers. Further research is needed to investigate the quantitative contribution of marine subsidies and the distance over which these are transferred into terrestrial systems.

In conclusion, our study shows that human transformation of coastal habitats has a substantial impact on the composition of the scavenging guild, and consequently on the speed with which marine carrion is removed from sandy beaches. As an increasing number of people live in expanding urban and peri-urban conurbations, it is important to develop a more robust understanding of how trends in human coastal demography and land-use affect key ecological processes and structures.

Acknowledgements

This project was supported through the Australian Research Council's Collaborative Research Network (CRN) programme and the Sunshine Coast Research Council – University of Sunshine Coast Strategic Partnership. We thank D. Chapman, M. Davis, D. Lynn and I. Schlacher for assistance in the field.

References

- Anderson, M. J. (2001). A new method for non-parametric multivariate analysis of variance? *Austral Ecology*, 26(1), 32–46.
- Barton, P. S., Cunningham, S. A., Lindenmayer, D. B., & Manning, A. D. (2013). The role of carrion in maintaining biodiversity and ecological processes in terrestrial ecosystems? *Oecologia*, 171(4), 761–772.
- Beasley, J. C., Olson, Z. H., & DeVault, T. L. (2012). Carrion cycling in food webs: Comparisons among terrestrial and marine ecosystems? *Oikos*, 121(7), 1021–1026.
- Berry, M. E., Bock, C. E., & Haire, S. L. (1998). Abundance of diurnal raptors on open space grasslands in an urbanized landscape. *Condor*, 100, 601–608.
- Blumstein, D. T., Fernandez-Juricic, E., Zollner, P. A., & Garity, S. C. (2005). Inter-specific variation in avian responses to human disturbance? *Journal of Applied Ecology*, 42(5), 943–953.
- Bodkin, J. L., & Jameson, R. J. (1991). Patterns of seabird and marine mammal carcass deposition along the central California coast, 1980–1986. *Canadian Journal of Zoology*, 69(5), 1149–1155.
- Bolker, B. M., Brooks, M. E., Clark, C. J., Geange, S. W., Poulsen, J. R., Stevens, M. H. H., et al. (2009). Generalized linear mixed models: A practical guide for ecology and evolution? *Trends in Ecology and Evolution*, 24(3), 127–135.
- Clergeau, P., Croci, S., Jokimäki, J., Kaisanlahti-Jokimäki, M. L., & Dinetti, M. (2006). Avifauna homogenisation by urbanisation: Analysis at different European latitudes? *Biological Conservation*, 127(3), 336–344.
- DeVault, T. L., Olson, Z. H., Beasley, J. C., & Rhodes, O. E. (2011). Mesopredators dominate competition for carrion in an agricultural landscape? *Basic and Applied Ecology*, 12(3), 268–274.
- DeVault, T. L., Rhodes, O. E., & Shivik, J. A. (2003). Scavenging by vertebrates: Behavioral, ecological, and evolutionary perspectives on an important energy transfer pathway in terrestrial ecosystems. *Oikos*, 102(2), 225–234.
- Duxbury, J., & Dickinson, S. (2007). Principles for sustainable governance of the coastal zone: In the context of coastal disasters? *Ecological Economics*, 63(2/3), 319–330.
- Eduardo, C., Carvalho, A., & Marini, M. A. N. (2007). Distribution patterns of diurnal raptors in open and forested habitats in south-eastern Brazil and the effects of urbanization? *Bird Conservation International*, 17(04), 367–380.
- Gotelli, N. J., & Colwell, R. K. (2001). Quantifying biodiversity: Procedures and pitfalls in the measurement and comparison of species richness? *Ecology Letters*, 4(4), 379–391.
- Gurran, N. (2008). The turning tide: Amenity migration in coastal Australia? *International Planning Studies*, 13(4), 391–414.
- Hamel, S., Killengreen, S. T., Henden, J. A., Eide, N. E., Roed-Eriksen, L., Ims, R. A., et al. (2012). Towards good practice guidance in using camera-traps in ecology: Influence of sampling design on validity of ecological inferences. *Methods in Ecology and Evolution*, 4(2), 105–113.
- Houston, D. C. (1986). Scavenging efficiency of turkey vultures in tropical forest. *Condor*, 88, 318–323.
- James, R. J. (2000). From beaches to beach environments: Linking the ecology, human-use and management of beaches in Australia. *Ocean and Coastal Management*, 43(6), 495–514.
- Longhurst, R. (1997). *Mooloolaba Surf Lifesaving Club: Our club our story 1922–1997*. Mooloolaba, QLD: The Mooloolaba Surf Lifesaving Club.
- Loss, S. R., Will, T., & Marra, P. P. (2013). The impact of free-ranging domestic cat on wildlife of the United States. *Nature Communications*, 4, 1396.

- Lotze, H. K., Lenihan, H. S., Bourque, B. J., Bradbury, R. H., Cooke, R. G., Kay, M. C., et al. (2006). Depletion, degradation, and recovery potential of estuaries and coastal seas. *Science*, 312(5781), 1806–1809.
- Lucrezi, S., Schlacher, T. A., & Walker, S. (2009). Monitoring human impacts on sandy shore ecosystems: A test of ghost crabs (*Ocypode* spp.) as biological indicators on an urban beach. *Environmental Monitoring and Assessment*, 152(1/4), 413–424.
- Maguire, G. S., Miller, K. K., Weston, M. A., & Young, K. (2011). Being beside the seaside: Beach use and preferences among coastal residents of south-eastern Australia? *Ocean and Coastal Management*, 54(10), 781–788.
- Maguire, G. S., Stojanovic, D., & Weston, M. A. (2010). Conditioned taste aversion reduces fox depredation on model eggs on beaches? *Wildlife Research*, 36(8), 702–708.
- Martínez, M. L., Intralawan, A., Vázquez, G., Pérez-Maqueo, O., Sutton, P., & Landgrave, R. (2007). The coasts of our world: Ecological, economic and social importance. *Ecological Economics*, 63(2/3), 254–272.
- McKinney, M. L. (2006). Urbanization as a major cause of biotic homogenization? *Biological Conservation*, 127(3), 247–260.
- McLachlan, A., & Brown, A. C. (2006). *The ecology of sandy shores*. Burlington, MA: Academic Press.
- Meager, J. J., Schlacher, T. A., & Nielsen, T. (2012). Humans alter habitat selection of birds on ocean-exposed sandy beaches? *Diversity and Distributions*, 18(3), 294–306.
- Micallef, A., & Williams, A. (2002). Theoretical strategy considerations for beach management? *Ocean and Coastal Management*, 45(4), 261–275.
- Morse, D. H. (1997). Distribution, movement, and activity patterns of an intertidal wolf spider *Pardosa lapidicina* population (Araneae, Lycosidae). *Journal of Arachnology*, 25(1), 1–10.
- Noriega, R., Schlacher, T. A., & Smeuninx, B. (2012). Reductions in ghost crab populations reflect urbanization of beaches and dunes? *Journal of Coastal Research*, 28(1), 123–131.
- Ogada, D., Torchin, M., Kinnaird, M., & Ezenwa, V. (2012). Effects of vulture declines on facultative scavengers and potential implications for mammalian disease transmission? *Conservation Biology*, 26(3), 453–460.
- Oksanen, J., Blanchet, F. G., Kindt, R., Legendre, P., Minchin, P. R., O'Hara, R. B., et al. (2012). vegan: Community Ecology Package. In *R package version 2.0-5*. <http://CRAN.R-project.org/package=vegan>
- Parmenter, R. R., & MacMahon, J. A. (2009). Carrion decomposition and nutrient cycling in a semiarid shrub-steppe ecosystem? *Ecological Monographs*, 79(4), 637–661.
- Peterson, C. A., Lee, S. L., & Elliott, J. E. (2001). Scavenging of waterfowl carcasses by birds in agricultural fields of British Columbia? *Journal of Field Ornithology*, 72(1), 150–159.
- Polis, G. A., & Hurd, S. D. (1996). Linking marine and terrestrial food webs: Allochthonous input from the ocean supports high secondary productivity on small islands and coastal land communities. *American Naturalist*, 147(3), 396–423.
- Polis, G. A., Sánchez-Piñero, F., Stapp, P. T., Anderson, W. B., & Rose, M. D. (2004). Trophic flows from water to land: Marine input affects food webs of islands and coastal ecosystems worldwide. In G. A. Polis, M. E. Power, & G. R. Huxel (Eds.), *Food webs at the landscape level* (pp. 200–216). Chicago: University of Chicago Press.
- Rassweiler, A., & Rassweiler, T. (2011). Does rapid scavenging hide non-predation mortality in coral-reef communities? *Marine and Freshwater Research*, 62(5), 510–515.
- Rose, M. D., & Polis, G. A. (1998). The distribution and abundance of coyotes: The effects of allochthonous food subsidies from the sea? *Ecology*, 79(3), 998–1007.
- Ruxton, G. D., & Houston, D. C. (2004). Obligate vertebrate scavengers must be large soaring fliers? *Journal of Theoretical Biology*, 228(3), 431–436.
- Schlacher, T. A., & Thompson, L. (2012). Beach recreation impacts benthic invertebrates on ocean-exposed sandy shores? *Biological Conservation*, 147(1), 123–132.
- Schlacher, T. A., Dugan, J., Schoeman, D. S., Lastra, M., Jones, A., Scapini, F., et al. (2007). Sandy beaches at the brink? *Diversity and Distributions*, 13(5), 556–560.
- Schlacher, T. A., Strydom, S., & Connolly, R. M. (2013). Multiple scavengers respond rapidly to pulsed carrion resources at the land–ocean interface. *Acta Oecologica*, 48, 7–12.
- Schlacher, T. A., Schoeman, D. S., Dugan, J., Lastra, M., Jones, A., Scapini, F., et al. (2008). Sandy beach ecosystems: Key features, sampling issues, management challenges and climate change impacts. *Marine Ecology*, 29, 70–90.
- Schlacher, T. A., Schoeman, D. S., Lastra, M., Jones, A., Dugan, J., Scapini, F., et al. (2006). Neglected ecosystems bear the brunt of change. *Ethology, Ecology and Evolution*, 18(4), 349–351.
- Selva, N., Jędrzejewska, B., Jędrzejewski, W., & Wajrak, A. (2005). Factors affecting carcass use by a guild of scavengers in European temperate woodland? *Canadian Journal of Zoology*, 83(12), 1590–1601.
- Short, J., & Smith, A. (1994). Mammal decline and recovery in Australia? *Journal of Mammalogy*, 75(2), 288–297.
- Spiller, D. A., Piovia-Scott, J., Wright, A. N., Yang, L. H., Takimoto, G., Schoener, T. W., et al. (2010). Marine subsidies have multiple effects on coastal food webs? *Ecology*, 91(5), 1424–1434.
- Tan, C. K. W., & Corlett, R. T. (2012). Scavenging of dead invertebrates along an urbanisation gradient in Singapore? *Insect Conservation and Diversity*, 5(2), 138–145.
- Warton, D. I., Wright, S. T., & Wang, Y. (2012). Distance-based multivariate analyses confound location and dispersion effects? *Methods in Ecology and Evolution*, 3(1), 89–101.
- Weston, M. A., Mcleod, E. M., Blumstein, D. T., & Guay, P. J. (2012). A review of flight initiation distances and their application to managing disturbance to Australian birds. *Emu*, 112, 269–286.
- Wilson, E. E., & Wolkovich, E. M. (2011). Scavenging: How carnivores and carrion structure communities? *Trends in Ecology and Evolution*, 26(3), 129–135.
- Wolcott, T. G. (1978). Ecological role of ghost crabs, *Ocypode quadrata* (Fabricius) on an ocean beach: Scavengers or predators? *Journal of Experimental Marine Biology and Ecology*, 31(1), 67–82.
- Zuur, A. F., Ieno, E. N., & Smith, G. M. (2007). *Analysing ecological data*. New York: Springer.